

## RESEARCH ARTICLE

## ASSESSMENT OF NATURAL RADIOACTIVITY LEVEL AND ASSOCIATED RADIOLOGICAL HAZARDS IN MARINE SEDIMENT SAMPLES COLLECTED FROM ABYAN BEACH, GULF OF ADEN, YEMEN

Fuad Abdo AS-Subaihi<sup>1,\*</sup>, Tahir Abdullah Abdulgabar Salem<sup>2</sup> and Mahmood Issa Ahmed<sup>3</sup><sup>1</sup> Dept. of Physics, Faculty of Education - Saber, University of Aden, Yemen<sup>2</sup> Dept. of Chemistry, Faculty of Education - Saber, University of Aden, Yemen<sup>3</sup> Dept. of Physics, Faculty of Education - Zanjibar, University of Aden, Yemen

\*Corresponding author: Fuad Abdo AS-Subaihi; E-mail: fabdo11@yahoo.com

Received: 20 December 2022 / Accepted: 07 January 2023 / Published online: 31 March 2023

## Abstract

Twenty three marine sediment samples taken from the side beach from the Abyan beach, Aden Gulf, Yemen to measuring the activity concentration of terrestrial radionuclides radium-226, thorium-232 and potassium-40 and its associated hazard indices using gamma spectrometer with High-Purity Germanium HPGe detector. The average activity concentration of natural radionuclides  $^{226}\text{Ra}$ ,  $^{232}\text{Th}$  and  $^{40}\text{K}$  in marine sediments samples under investigation were  $16.25 \pm 0.63 \text{ Bq kg}^{-1}$ , for Ra-226,  $23.80 \pm 1.46 \text{ Bq kg}^{-1}$  for Th-232 and  $518.54 \pm 44.61 \text{ Bq kg}^{-1}$  for K-40 respectively. The radiation hazard indices which resulting from the presence of natural radionuclides in marine sediment samples were calculated and the obtained results indicate that the average value of radium equivalent activity was  $90.21 \text{ Bq kg}^{-1}$ . The average value of External hazard index  $H_{\text{ex}}$  was 0.243 and the average values of internal hazard index  $H_{\text{in}}$  was 0.287. The average value of outdoor absorbed dose rate  $D_{\text{out}}$  was  $43.66 \text{ nGy h}^{-1}$  and indoor absorbed dose rate  $D_{\text{in}}$  was  $83.13 \text{ nGy h}^{-1}$ . The average values of annual outdoor effective dose  $\text{AED}_{\text{out}}$  was  $0.053 \text{ mSv y}^{-1}$  and the annual indoor effective dose  $\text{AED}_{\text{in}}$  was  $0.407 \text{ mSv y}^{-1}$ . The average values of the annual gonadal dose equivalent  $\text{AGDE}$  were  $312.52 \mu\text{Sv y}^{-1}$ . While the average values of outdoor excess life-time cancer risk  $\text{ELCR}_{\text{out}}$  was  $0.187 \times 10^{-3}$  and the indoor excess life-time cancer risk  $\text{ELCR}_{\text{in}}$  was  $1.427 \times 10^{-3}$ . They are less than worldwide limits.

**Keywords:** Natural radioactivity, Abyan beach, Radiation hazards, Marine sediments, HPGe detector.

## 1. Introduction

Human environment is naturally radioactive, and human beings are exposed to radiation arising from natural sources, including cosmic and terrestrial origin, in addition to artificial radioactivity from fallout in nuclear testing and medical applications. Natural sources contribute approximately 80% of the environmental radiation [1]. Radioactivity naturally exists in the environment in different conditions such as soil, underground water, marine sediment, and biota. Radioactivity enters the marine environment through different pathways, including via river and rainwater transport into the sea; however, this is often due to nuclear waste disposal, which is discharged from nuclear power plants as well as from medical, industrial, research, and educational uses of radionuclides [2]. The activity concentration levels of terrestrial radioactive nuclides which found in air, soil, underground water, marine sediment, and biota and other component of the

environment are depending on the properties of the geological, geochemical and geographical of the region under studied and appear at different rates of the world [3, 4]. There are few areas in the world such as Brazil, China, India, Austria, France and Iran where the background radiation levels were found to be high; varying over an order of magnitude depending upon the site-specific terrestrial radioactivity. [5, 6, 7]. Radiation and radioactivity in the environment have natural and man-made sources. Exposure to natural radiation represents the most significant part of the total exposure to radiation in the environment [8, 9, 10]. Only natural radionuclides with half-lives comparable with the age of the earth or their corresponding decay products existing in terrestrial material such as  $^{232}\text{Th}$ ,  $^{238}\text{U}$ ,  $^{235}\text{U}$ ,  $^{226}\text{Ra}$ ,  $^{228}\text{Ra}$  and  $^{40}\text{K}$  are of great interest. The levels of these radionuclides are relatively distributed in soil based on the nature of its geological formations [10, 11]. The uptake of radionuclides by marine sediments depends on

their physical and chemical properties. The radionuclides distribution in marine sediments provides essential information concerning sediments movement and accumulation that provide a strong signal indicating sediment origin [12, 13]. Assessing radiation exposure among humans requires a better understanding of the radionuclide's behavior in pertinent environments. Thus, the primary aims of nearly all marine radioactivity studies have been to form a scientific foundation upon which to determine the radiological risk of radioisotopes in marine environments. This is an enormously important issue that is in alignment with the present radiation protection Standards [14, 15, 16].

An attempt is made in this paper to determine the activity concentration of  $^{226}\text{Ra}$ ,  $^{232}\text{Th}$  and  $^{40}\text{K}$  in marine sediment samples collected from Abian beach, Aden Gulf, Yemen using high purity germanium (HPGe) detector and to compute the total absorbed gamma-dose rate in air due to the presence of  $^{238}\text{U}$  series,  $^{232}\text{Th}$  series and  $^{40}\text{K}$  in the samples. It is necessary to estimate the doses received and then compare such data with the nearest relevant data for reference organisms to evaluate the likely radiation effects for such organisms in an environmental context [17]. The aim of the present work to provides database

about radium-226, thorium-232 and potassium-40 concentration in marine sediments of Abyan beach, Aden Gulf, Yemen. This data can be used as a reference data for future studies and it may be useful for complete radioactivity mapping for Yemen republic.

## 2. Materials and Methods

### 2.1. Sampling and samples preparation:

Twenty three sand beach sediment samples have been collected from predetermined undisturbed areas alongside the coast of Abyan beach, Aden Gulf, Yemen during low-tide (figure.1). At each of the sample sites, a sampling area of  $1\text{m}^2$  was considered, where five wet samples were taken, each weighing about 1kg. Four samples were taken from the corners of the site and the fifth was from the center. The sediment samples were taken at depth of 5cm from the surface; the samples were collected using Grab sample equipment. Then separated from the contamination materials and air-dried at room temperature for a week, then dried to  $110^\circ\text{C}$ , milled and sieved through 0.2 mm sieve. The dried samples were put inside cylindrical can. The cans were sealed, gas-tight and stored for four weeks for secular equilibrium [4, 18].

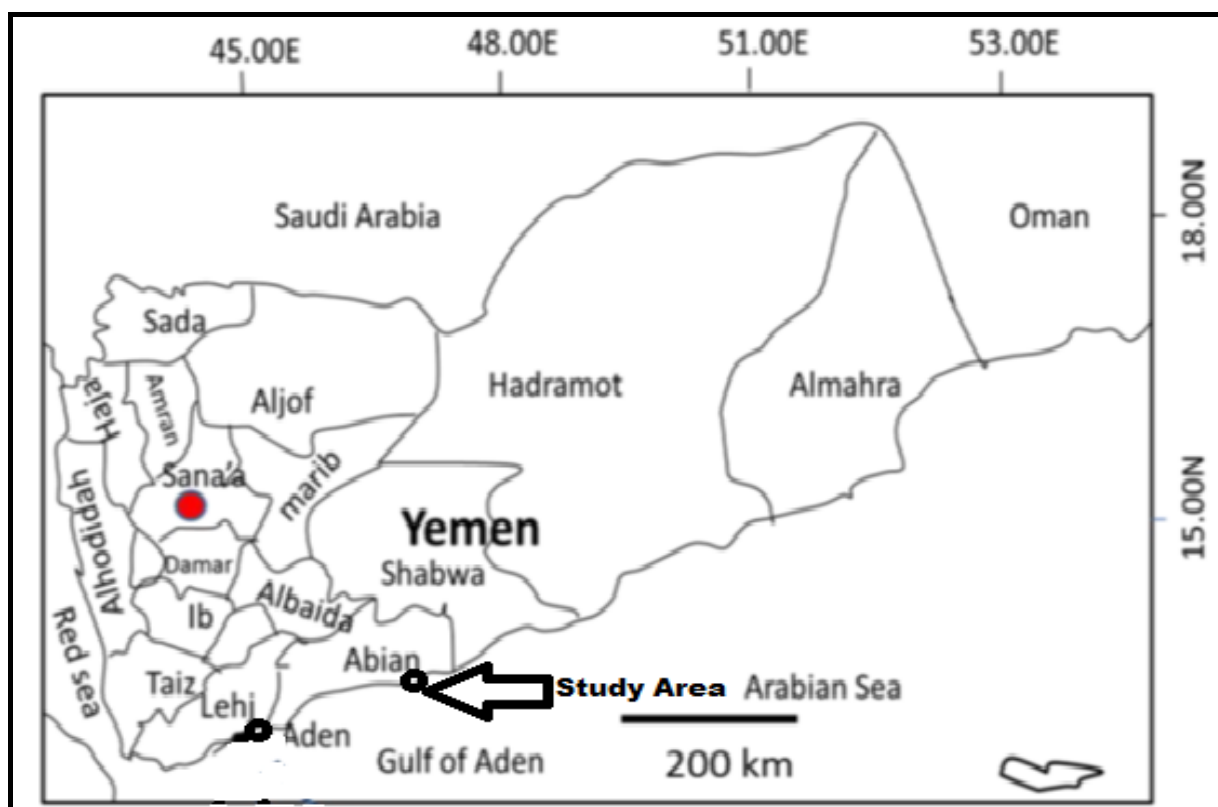


Fig. 1: Map of the study area.

## 2.2. Experimental Setup

The measurement of the activity concentration of the  $^{238}\text{U}$ ,  $^{232}\text{Th}$  was by way of the daughter products. For a nuclide having more than one peak in the spectrum, the activity concentration was obtained as the weighted average activity at each peak. The emissions of  $^{214}\text{Bi}$  (609.31, 1120.29, 1764.49keV) and  $^{214}\text{Pb}$  (295.22, 351.93keV) were used to determine the activity of  $^{238}\text{U}$ . The activity of  $^{232}\text{Th}$  was estimated from the  $^{212}\text{Bi}$ ,  $^{212}\text{Pb}$ , and  $^{228}\text{Ac}$  radionuclide activities measured directly from their gamma-ray energy lines 727.17, 238.63, and 911.60keV, respectively. The activity concentrations of  $^{40}\text{K}$  were determined using its only  $\gamma$ - ray line of peak energy 1460.82keV. Prior to sample measurement, the background was determined with an empty Marinelli beaker under identical measurement conditions as the samples. Counting time was 72, 0 0 0 s. The data acquisition, display and on-line spectrum analysis were carried out using the Genie 2000, spectroscopy software from Canberra [7]. The activity concentration ( $A_{\text{Ei}}$ ) (Bq/kg) of each radionuclide in any given sample was calculated from the spectrum using the following analytical expression:

$$A_{\text{Ei}} = \frac{NP}{t_c \cdot I_{\gamma}(E_{\gamma}) \cdot \varepsilon(E_{\gamma}) \cdot M} \quad (1)$$

Where NP is the number of count in a given peak area corrected for background peaks of a peak at energy E,  $\varepsilon(E_{\gamma})$  the detection efficiency at energy E,  $t_c$  is the counting life time,  $I_{\gamma}(E_{\gamma})$  the number of gammas per disintegration of this nuclide for a transition at energy E, and M the mass in kg of the measured sample [15, 16].

## 3. Results and Discussion

### 3.1. Activity concentrations in Coastal Marine Sediment Samples

The activity concentration in (Bq/kg) of  $^{226}\text{Ra}$ ,  $^{232}\text{Th}$  and  $^{40}\text{K}$  for Coastal Marine sediment samples collected from the coast of Abian beach, Aden Gulf, Yemen are listed in (Table1) and shown in figure (2). The mass concentrations of Th and U were calculated from  $^{238}\text{U}$  (assuming secular equilibrium between  $^{238}\text{U}$  and  $^{226}\text{Ra}$ , as well as their progenies) and from  $^{232}\text{Th}$  (assuming secular equilibrium between  $^{232}\text{Th}$  and  $^{228}\text{Ra}$ ). The activity concentrations of  $^{226}\text{Ra}$ ,  $^{232}\text{Th}$ , and  $^{40}\text{K}$  in Coastal Marine Sediment samples were found to vary between  $9.44 \pm 0.38$  to  $36.30 \pm 1.41 \text{ Bqkg}^{-1}$  with average value  $16.25 \pm 0.63 \text{ Bqkg}^{-1}$  for  $^{226}\text{Ra}$ . The lowest level was observed at sample number 23 and the highest was found at sample number 11. It is clear that the highest level of  $^{226}\text{Ra}$  is 4-fold of the lowest level.  $^{232}\text{Th}$  results indicated a range from  $12.94 \pm 0.79$  to  $95.75 \pm 5.88 \text{ Bqkg}^{-1}$  with

average value  $23.80 \pm 1.46 \text{ Bqkg}^{-1}$ . The lowest value was recorded at sample number 12 and the highest levels were at sample number 11. The range of  $^{232}\text{Th}$  showed that the highest level is 7-fold of the lowest value. The difference between the sample in the west and the sample in the east could be explained by the presence of black sands deposits, which contain heavy minerals such as monazite. These sands contain, overall, two orders of magnitude more of  $^{232}\text{Th}$  and  $^{238}\text{U}$  decay series radionuclides [19, 12].  $^{40}\text{K}$  results indicated a range from  $296.67 \pm 25.52 \text{ Bqkg}^{-1}$  to  $747.72 \pm 64.32 \text{ Bqkg}^{-1}$  with the average value  $518.54 \pm 44.61 \text{ Bqkg}^{-1}$ . The lowest level was observed at sample number 11 and the highest was found at sample number 05. The bar diagrams distributions of radioactivity levels over the investigated samples are shown in Figure (2) for  $^{226}\text{Ra}$ ,  $^{232}\text{Th}$  and  $^{40}\text{K}$ , respectively. The worldwide concentrations of the radionuclides U-238, Th-232, and K-40 have averages in marine sediment samples of 35, 30, and  $400 \text{ Bqkg}^{-1}$  respectively (UNSCEAR, 1993) [2]. Our results show that the mean activity concentrations of  $^{226}\text{Ra}$  and  $^{232}\text{Th}$  radionuclides are lower than the worldwide concentrations according to the UNSCEAR report [2, 20]. These average values agreed well with the result obtained in earlier studies for  $^{226}\text{Ra}$  and  $^{232}\text{Th}$  [12]. The average activity concentrations of  $^{40}\text{K}$  for all samples were higher than the world average value  $400 \text{ Bqkg}^{-1}$  [4]. A large variability in activity concentrations is shown among radionuclides, reflecting the geological and morphological characteristics of the collected sediments, as well as their respective radionuclide contents. A high degree of variability in the measured radioactivity was shown in the studied samples, as these samples reflect the geological characteristics of their sites of origin. Usually, the radioactivity of  $^{238}\text{U}$  and  $^{232}\text{Th}$  is linked with heavy minerals, while that of  $^{40}\text{K}$  is associated with clay minerals. In Table 2, a comparison is given of the average (range) radioactivity concentrations ( $\text{Bqkg}^{-1}$ ) obtained in this work versus in the literature. According to the IAEA, when the activity of the  $^{238}\text{U}$  or  $^{232}\text{Th}$  decay series is  $\leq 1000 \text{ Bqkg}^{-1}$  and that of  $^{40}\text{K}$  is  $\leq 10,000 \text{ Bqkg}^{-1}$ , the radioactive material may not be regarded as naturally occurring and is thus exempt from regulations [8, 16].

**Table 1:** Activity concentration (Bq.Kg<sup>-1</sup>) of <sup>238</sup>U, <sup>232</sup>Th and <sup>40</sup>K in Coastal Marine Sediment samples.

Sample Number	Coordinates		Activity concentration (Bq/ kg)		
	Latitude(N)	Longitude (E)	<sup>226</sup> Ra	<sup>232</sup> Th	<sup>40</sup> K
S 01	12°53'36.1"	44°54'32.0"	15.54±0.61	19.31±1.22	447.07±38.458
S 02	12°56'80.9"	45°06'12.9"	17.86±0.73	29.73±1.82	588.09±50.589
S 03	12°57'31.4"	45°06'31.9"	26.47±1.02	34.52±2.11	556.55±47.876
S 04	12°57' 58.9"	45°06'72.3"	15.83±0.61	15.87±0.96	725.27±62.39
S 05	12°58'03.0"	45°07'07.0"	21.63±0.84	20.89±1.28	747.72±64.32
S 06	12°53'61.9"	45°07'69.4"	30.40±1.17	37.33±2.29	632.75±54.43
S 07	12°59'08.0"	45°08'29.6"	14.65±0.56	18.59±1.14	604.01±51.959
S 08	12°59'48.7"	45°08'93.8"	14.17±0.55	16.27±0.99	658.26±56.62
S 09	12°59'71.7"	45°09'36.9"	17.84±0.69	21.20±1.29	622.86±53.58
S 10	12°59'97.6"	45°09'90.6"	11.44±0.44	13.29±0.81	584.06±50.24
S 11	13°00'37.8"	45°10'82.3"	36.30±1.41	95.75±5.88	296.67±25.52
S 12	13°00'80.2"	45°11'90.7"	10.70±0.41	12.94±0.79	478.49±41.16
S 13	13°01'17.5"	45°12'93.9"	12.15±0.49	13.55±0.82	585.13±50.33
S 14	13°01'47.8"	45°13'89.7"	12.18±0.47	17.00±1.03	485.73±41.78
S 15	13°01'78.6"	45°14'67.4"	11.89±0.46	13.61±0.83	554.15±47.67
S 16	13°01'.980"	45°15'.591"	13.07±0.51	17.65±1.08	457.33±39.34
S 17	13°02'13.4"	45°16'20.0"	11.80±0.45	14.97±0.92	533.57±45.899
S 18	13°04'63.2"	45°24'22.7"	14.24±0.55	22.45±1.37	301.96±25.97
S 19	13°04'00"	45°23'.042"	11.57±0.45	14.31±0.87	586.37±50.44
S 20	13°03'19.7"	45°22'03.1"	15.74±0.62	34.52±2.11	326.53±28.089
S 21	13°02'.506"	45°17'84.1"	13.45±0.52	20.13±1.24	349.71±30.08
S 22	13°02'42.8"	45°17'43.3"	15.43±0.59	24.10±1.48	434.70±37.39
S 23	13°02'21.7"	45°16'52.2"	9.44±0.38	19.36±1.19	369.51±31.786
Minimum			9.44±0.38	12.94±0.79	296.67± 25.52
Maximum			36.30±1.41	95.75±5.88	747.72± 64.32
Mean			16.25± 0.63	23.80± 1.46	518.54± 45.84

**Table 2:** Comparison of average (range) radioactivity concentrations (Bqkg<sup>-1</sup>) with other studies.

Country	<sup>226</sup> Ra	<sup>232</sup> Th	<sup>40</sup> K	Reference
Worldwide	35	30	400	[20]
Abian beach, Aden Gulf, Yemen	9.44-36.30	12.94-95.75	296.67-747.72	Present study
Gulf of Oman	9.3 - 24.8	10.4 - 54.9	29.0–78.7	[16]
Arab Sea Qatar	4.2–19.5	1.0–6.0	11–188	[21]
Arab Sea Kuwait	17.3–20.5	15–16.4	353– 445	[22]
Arab Sea Kuwait	18.6–21.4	14.0–17.1	351.2– 404	[23]
Saudi Arabia	4.4–19.3	5.3–58.9	324.6–1133	[24]
Red Sea, Egypt	1.92- 17.55	5.62- 28.77	123.27- 277.38	[4]
Alexandria beach, Egypt	12.6-499.18	0.65–386.2	10-122	[12]
Kuakata beach, Bangladesh	48.76 ±5.29	126.11 ±3.31	292.38 ±18.24	[6]
Rizhao beaches (China)	12	15	1079	[25]
Beaches of Aegean sea (Turkey)	290	1160	532	[26]
Sediments of Cadiz Bay (Spain)	13	451	19	[27]
Beach of Jeddah Saudi Arabia	14.22	14.00	968.19	[28]
Sediment, Saudi Arabia	18.3-37.6	7.8-25.5	202-432	[29]
Surface Sediment Yangtze Estuary	13.7–52.	26.1–71.9	392–898	[30]

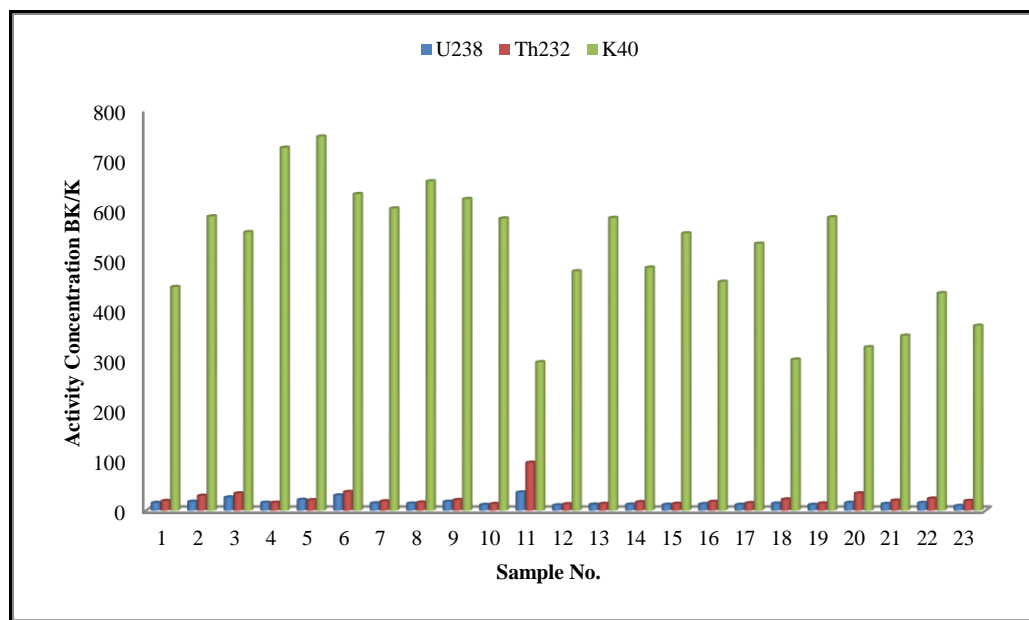


Fig.2:  $^{226}\text{Ra}$ ,  $^{232}\text{Th}$ , and  $^{40}\text{K}$  Activity concentration ( $\text{Bq.Kg}^{-1}$ ) in Marine sediment sample.

### 3.2 Assessing Radiological Hazards

#### 3.2.1. Radium equivalent activity ( $R_{eq}$ )

Radium-equivalent activity ( $R_{eq}$ ) is a single parameter that represents the collective risk of  $^{226}\text{Ra}$ ,  $^{232}\text{Th}$ , and  $^{40}\text{K}$  radioactivity [16, 31, 32]. This parameter can be used to assess whether external doses to the public exceed the recommended annual dose limit of  $1\text{mSv}$ .  $R_{eq}$  was calculated from the next equation (2):

$$R_{eq} (\text{Bq.Kg}^{-1}) = A_{Ra} + 1.43A_{Th} + 0.077A_K \quad (2)$$

Where  $A_{Ra}$ ,  $A_{Th}$ , and  $A_K$ , are the activity concentrations of  $^{226}\text{Ra}$ ,  $^{232}\text{Th}$  and  $^{40}\text{K}$  respectively. Table 3 presented the measured radium equivalent activity  $R_{eq}$  in sediment samples which were ranged from  $65.58$  to  $196.06 \text{Bqkg}^{-1}$  with an average value  $90.21 \text{Bqkg}^{-1}$ . The results are lower than the recommended limit  $370 \text{Bqkg}^{-1}$  reported by UNSCEAR (1988,2000) [1, 33]. The behavior of radium equivalent activity  $R_{eq}$  is shown in Figure 3.

#### 3.2.2. External Hazard Index ( $H_{ex}$ )

The external radiation exposure due to natural radioactivity is defined in terms of the external hazard index ( $H_{ex}$ ), calculated as follows [16, 2, 31]:

$$H_{ex} = (A_{Ra}/370 + A_{Th}/259 + A_K/4810) \leq 1 \quad (3)$$

Where  $A_{Ra}$ ,  $A_{Th}$ , and  $A_K$  are the specific activities ( $\text{Bqkg}^{-1}$ ) of  $^{226}\text{Ra}$ ,  $^{232}\text{Th}$ , and  $^{40}\text{K}$ , respectively in the studied samples. To comply with the requirements of the  $1\text{mSv}$  annual dose limit for the public,  $H_{ex}$  should be  $<1$ , as shown above [20]. As seen in Table 3, the  $H_{ex}$  values ranged from  $0.177$  to  $0.529$  with an average value ( $0.24$ ). These results ensure that the public's exposure to the environmental radioactivity of  $^{226}\text{Ra}$ ,  $^{232}\text{Th}$ , and  $^{40}\text{K}$  radionuclides in beach sediment remain within acceptable limits.

#### 3.2.3. Internal hazards index ( $H_{in}$ )

In addition to the external radiation hazard they pose, radon and its short-lived daughters are also hazardous to the respiratory organs. The internal exposure caused by radon and its daughter products is quantified by the internal hazard index  $H_{in}$ , which has been calculated as follows [16, 2, 31]:

$$H_{in} = (A_{Ra}/185 + A_{Th}/259 + A_K/4810) \leq 1 \quad (4)$$

The values of internal hazard index  $H_{in}$  As seen in Table 3 ranged from  $0.203$  to  $0.627$  with an average value  $0.29$ , which is lower than the worldwide limit  $\leq 1$  recommended by UNSCEAR (2000) [33]. The behavior of external hazard index ( $H_{ex}$ ) and internal hazard index  $H_{in}$  is shown in Figure 4.

#### 3.3. Absorbed dose rate

Absorbed dose ( $D$ ) assess the energy which stored in a medium due to the ionizing radiation emitted from natural occurring radionuclides  $^{238}\text{U}$ ,  $^{232}\text{Th}$ , and  $^{40}\text{K}$  and measured in SI units as joules per kilogram Gray (Gy). It is calculated based on guide lines provided by UNSCEAR 2000 [33]. To convert the activity concentration of  $^{238}\text{U}$ ,  $^{232}\text{Th}$ , and  $^{40}\text{K}$  radionuclides into doses in ( $\text{nGyh}^{-1} \text{per Bqkg}^{-1}$ ), UNSCEAR (1988) has given the dose conversion factors as  $0.427$ ,  $0.662$  and  $0.043$  respectively [1]. Using these factors, the outdoor absorbed dose rate is calculated using the following equation [2]:

$$D_{out} (\text{nGy.h}^{-1}) = 0.462A_{Ra} + 0.604A_{Th} + 0.042A_K \quad (5)$$

The indoor gamma ray dose imparted by of  $^{238}\text{U}$ ,  $^{232}\text{Th}$ , and  $^{40}\text{K}$  radionuclides present indoor can be calculated by converting the absorbed dose rate to effective dose by using the conversion factors  $0.92$ ,  $1.1$  and  $0.081 \text{nGyh}^{-1} \text{per Bqkg}^{-1}$  respectively. By using these



factors following, the following equation is used to calculate the indoor dose rate, given by UC European Commission (1999) [34]:

$$D_{in} (nGy.h^{-1}) = 0.92A_U + 1.1A_{Th} + 0.081A_K \quad (6)$$

Table 3 shows the calculated outdoor absorbed dose  $D_{out}$  due to the presence of  $^{238}U$ ,  $^{232}Th$ , and  $^{40}K$  in the marine sediments of Abian beach, that ranged from 31.57 to 87.06 nGy $h^{-1}$  with an average value of 43.66 nGy $h^{-1}$  which is lower than the worldwide limit of 59 nGy $h^{-1}$  as presented by UNSCEAR (2000) [33]. The values of

indoor absorbed dose  $D_{in}$  calculated during present study range from 59.91 to 162.75 nGy $h^{-1}$  with an average of 83.13 nGy $h^{-1}$ . Our results are higher than the doses reported in Gulf of Oman (49.26 nGy $h^{-1}$ ) [16]. Our results are within the worldwide limit of 84 nGy $h^{-1}$  as recommended by UNSCEAR (2000) [33]. The behavior of indoor and outdoor absorbed dose in samples is shown in Figure 5.

**Table 3:** Radium-equivalent activity, absorbed dose rates, effective rates, excessive cancer risk, annual gonadal dose and external hazard index ( $H_{ex}$ ) associated with the radioactivity in Coastal Marine Sediment samples.

Samples code	$Ra_{eq}$ (Bq/kg)	$D_{out}$ (nGy/h)	$D_{in}$ (nGy/h)	$AEDE_{out}$ ( $\mu Sv/y$ )	$AEDE_{in}$ (mSv/y)	$AGDE$ ( $\mu Sv/y$ )	$H_{ex}$	$H_{in}$	$ELCR_{in}$	$ELCR_{out}$
S 01	77.58	37.62	71.75	0.046	0.352	269.11	0.209	0.251	1.23	0.161
S 02	105.66	50.91	96.77	0.062	0.475	364.13	0.285	0.333	1.66	0.218
S 03	118.68	56.45	107.39	0.069	0.527	400.82	0.320	0.392	1.84	0.242
S 04	94.37	47.36	90.76	0.058	0.445	342.98	0.255	0.297	1.56	0.203
S 05	109.08	54.02	103.44	0.066	0.507	388.95	0.294	0.353	1.78	0.232
S 06	132.49	63.16	120.28	0.077	0.590	448.65	0.358	0.439	2.06	0.271
S 07	87.75	43.37	82.85	0.053	0.406	312.64	0.237	0.276	1.42	0.186
S 08	88.12	44.02	84.25	0.053	0.413	318.48	0.238	0.276	1.45	0.189
S 09	96.12	47.21	90.19	0.058	0.442	339.33	0.259	0.308	1.55	0.203
S 10	75.42	37.84	72.46	0.046	0.355	274.31	0.204	0.234	1.24	0.162
S 11	196.06	87.06	162.75	0.107	0.798	605.55	0.529	0.627	2.79	0.374
S 12	66.05	32.86	62.84	0.040	0.308	237.40	0.178	0.207	1.08	0.141
S 13	76.58	38.37	73.48	0.047	0.360	277.92	0.207	0.239	1.26	0.165
S 14	73.89	36.29	69.25	0.044	0.339	261.23	0.199	0.232	1.19	0.156
S 15	74.023	36.99	70.79	0.045	0.347	267.64	0.199	0.232	1.21	0.159
S 16	73.53	35.91	68.48	0.044	0.336	257.77	0.198	0.234	1.17	0.154
S 17	74.29	36.90	70.54	0.045	0.346	266.57	0.201	0.232	1.21	0.158
S 18	69.59	32.82	62.26	0.040	0.305	232.66	0.188	0.226	1.07	0.141
S 19	77.18	38.62	73.88	0.047	0.362	279.69	0.208	0.239	1.27	0.166
S 20	90.25	41.84	78.90	0.051	0.387	295.47	0.244	0.286	1.35	0.179
S 21	69.17	33.06	62.85	0.040	0.308	235.54	0.187	0.223	1.08	0.142
S 22	83.36	39.94	75.91	0.049	0.372	284.91	0.225	0.267	1.30	0.171
S 23	65.58	31.57	59.91	0.039	0.294	226.13	0.177	0.203	1.029	0.135
Worldwide	370	59	84	0.07	0.45	100	$\leq 1$	$\leq 1$	1.16	1.45
Min.	65.58	31.57	59.91	0.039	0.294	226.13	0.177	0.203	1.029	0.135
Max.	196.06	87.06	162.75	0.107	0.798	605.55	0.529	0.627	2.79	0.374
Mean	90.21	43.66	83.13	0.053	0.407	312.52	0.243	0.287	1.427	0.187

$Ra_{eq}$  is the radium equivalent activity;  $D_{in}$  and  $D_{out}$  are the indoor and outdoor air absorbed, respectively.

$AED_{Total}$  is the total effective doses due to internal and external radiation exposure.  $ECR$ , excessive cancer risk.

$AGD$  (mGy.y $^{-1}$ ), annual gonadal dose and  $H_{ex}$  is the external hazard index.

### 3.4. Annual effective dose equivalent (*AEDE*)

The outdoor annual effective dose equivalent (*AEDE<sub>out</sub>*) was estimated from outdoor absorbed dose in air to convert it to effective dose. While the indoor annual effective dose equivalent (*AEDE<sub>in</sub>*) is estimated from indoor absorbed dose in air to convert it to the effective dose. UNSCEAR (2000) [33] reported the value  $0.7\text{SvGy}^{-1}$  as conversion coefficient from absorbed dose in the air to the effective dose received by adults. While 0.2 and 0.8 represent the outdoor and indoor occupancy factors respectively. The annual indoor and outdoor effective dose rate in  $\text{mSvy}^{-1}$  is given by the following formulas respectively as reported by UNSCEAR (2000) [33]:

$$\text{AEDE (mSvy}^{-1}\text{)} = \text{D (nGyh}^{-1}\text{)} \times 8760\text{h} \times 0.8 \times 0.7 \text{ SvGy}^{-1} \times 10^{-6} \quad (7)$$

$$\text{AEDE (mSvy}^{-1}\text{)} = \text{D (nGyh}^{-1}\text{)} \times 8760\text{h} \times 0.2 \times 0.7 \text{ SvGy}^{-1} \times 10^{-6} \quad (8)$$

The values of outdoor annual effective dose equivalent *AEDE<sub>out</sub>* for Abyen beach sediments ranges from  $0.039$  to  $0.107\text{mSvy}^{-1}$  with an average of  $0.053\text{mSvy}^{-1}$  which is lower than the worldwide average of  $0.07\text{mSvy}^{-1}$ .

The indoor annual effective dose equivalent *AEDE<sub>in</sub>* calculated with a range from  $0.29$  to  $0.798\text{mSvy}^{-1}$  with an average value  $0.41\text{mSvy}^{-1}$  which is same as the worldwide limit of  $0.41\text{mSvy}^{-1}$  reported by UNSCEAR (2000) [33]. The total average annual effective dose equivalent was estimated to be  $0.44\text{mSvy}^{-1}$  which is within the worldwide limit of  $0.48\text{mSv/y}$  as predicted by UNSCEAR (2000) [34]. Spatial distribution of total annual effective dose equivalent in studied area is shown in Figure 6. From Table 3 it can be seen that the estimated values of annual effective dose obtained in this study is higher than the value reported in Gulf of Oman ( $0.27\text{mSvy}^{-1}$ ) [16].

### 3.5. Annual Gonadal Dose Equivalent (*AGDE*)

The United Nations Scientific Committee on the Effects of Atomic Radiation UNSCEAR (1988) [1] has been interested active bone marrow and bone surface cells as organs. Therefore, the annual gonadal dose equivalent (*AGDE*) due to the specific activities of  $^{226}\text{Ra}$ ,  $^{232}\text{Th}$ , and  $^{40}\text{K}$  in studied samples was estimated using the following equation: [33].

$$\text{AGDE (}\mu\text{Sv/y)} = 3.09A_U + 4.18A_{Th} + 0.314A_K \quad (9)$$

The obtained values of *AGDE* are listed in Table 3. The values of *AGDE* varied from  $226.13$  to  $605.55\mu\text{Svy}^{-1}$  and the average value was found to be  $312.52\mu\text{Svy}^{-1}$ . The average value of *AGDE* was found to be  $181.1\mu\text{Svy}^{-1}$  in Gulf of Oman [16]. Also, the average value of *AGDE* was found to be  $2850\mu\text{Svy}^{-1}$  in Kerala, India [35]. These two values of *AGDE* are higher than our result in Kerala, India and lower than our result in Gulf of Oman. The obtained *AGDE* value for marine beach sediments in Abian beach is higher than the worldwide limit of  $300\mu\text{Svy}^{-1}$  according to UNSCEAR reports [2]. The behavior of *AGDE* in samples is shown in Figure 7.

### 3.6. Life-time cancer risk (*ELCR*)

The excess life-time cancer risk (*ELCR*) was estimated from annual effective dose equivalent using the equation [36, 37]:

$$\text{ELCR}_{out} = \text{AEDE}_{out} \times DL \times RF \quad (10)$$

$$\text{ELCR}_{in} = \text{AEDE}_{in} \times DL \times RF \quad (11)$$

Where  $DL$ , and  $RF$  are the duration of life (70 years), and risk factor ( $0.05/\text{Sv}$ ), respectively. Defined the risk factor as fatal cancer risk per Sievert is assigned a value of  $0.05$  by ICRP (2012) [37] for the public for random effects, for low-level radiations.

The excess lifetime cancer risk *ELCR* for outdoor exposure calculated for sediment, given in Table 3 varied between  $0.135 \times 10^{-3}$  and  $0.374 \times 10^{-3}$  with average value of  $0.187 \times 10^{-3}$ . This value was found to be less than the limit of  $0.29 \times 10^{-3}$  set by UNSCEAR (2000) [33].

For indoor exposure it is varied from  $1.029 \times 10^{-3}$  to  $2.79 \times 10^{-3}$  with an average of  $1.427 \times 10^{-3}$ . It is higher than the limit  $1.16 \times 10^{-3}$  [36]. The total *ELCR* ranges from  $1.16 \times 10^{-3}$  to  $3.17 \times 10^{-3}$  with an average value of  $1.61 \times 10^{-3}$ . It clear that the *ELCR* for marine beach sediments in Abian beach is higher than the worldwide limit of  $1.45 \times 10^{-3}$  [33]. The behavior of *ELCR* in samples is shown in Figure 8.

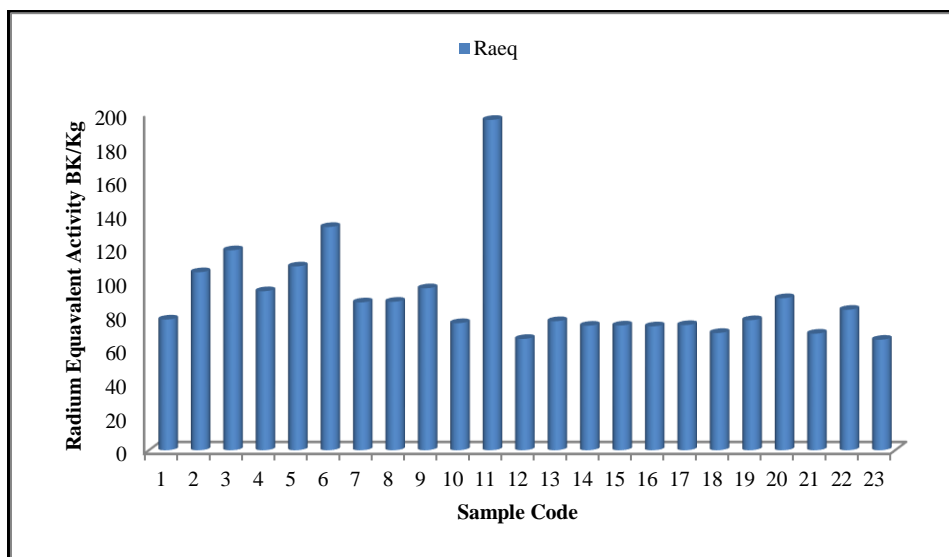


Fig. 3: Radium Equivalent Activity of sediment sample

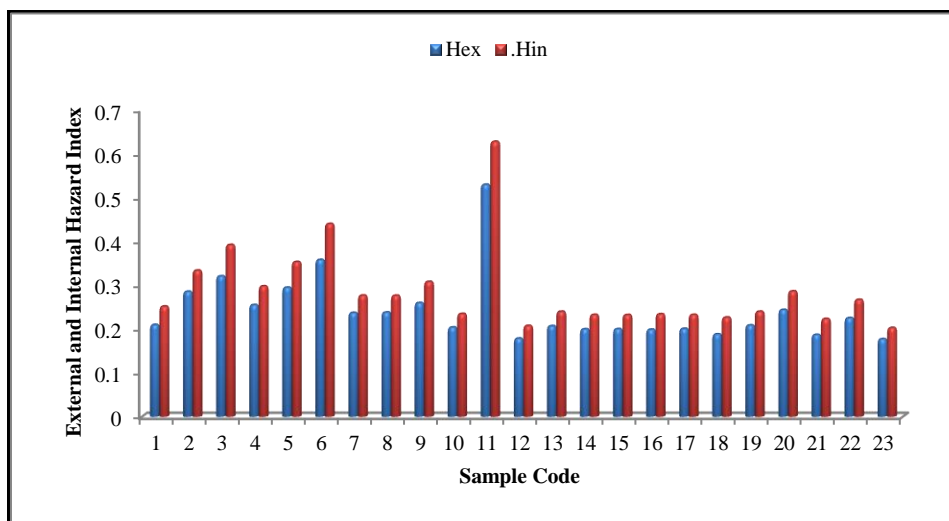


Fig. 4: Hazard Index of sediment samples.

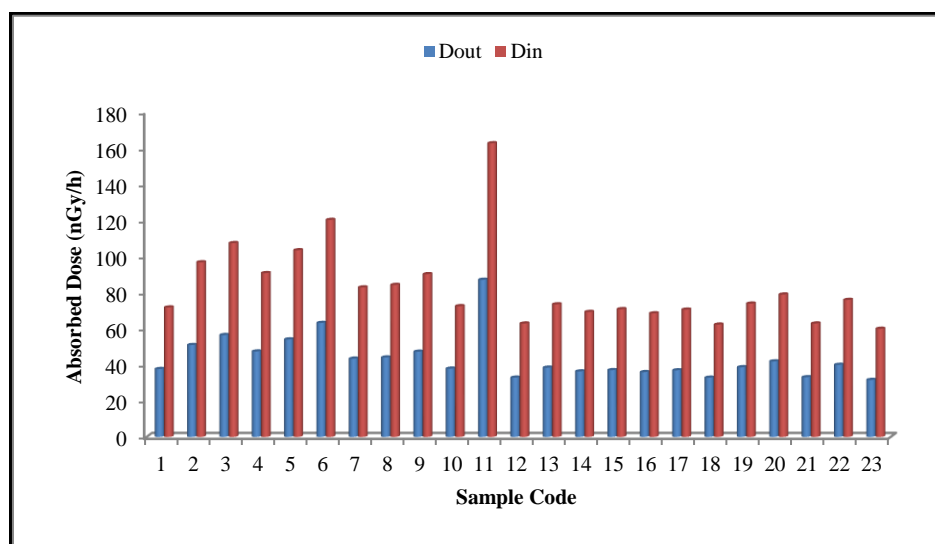


Fig. 5: Absorbed Dose of sediment samples.



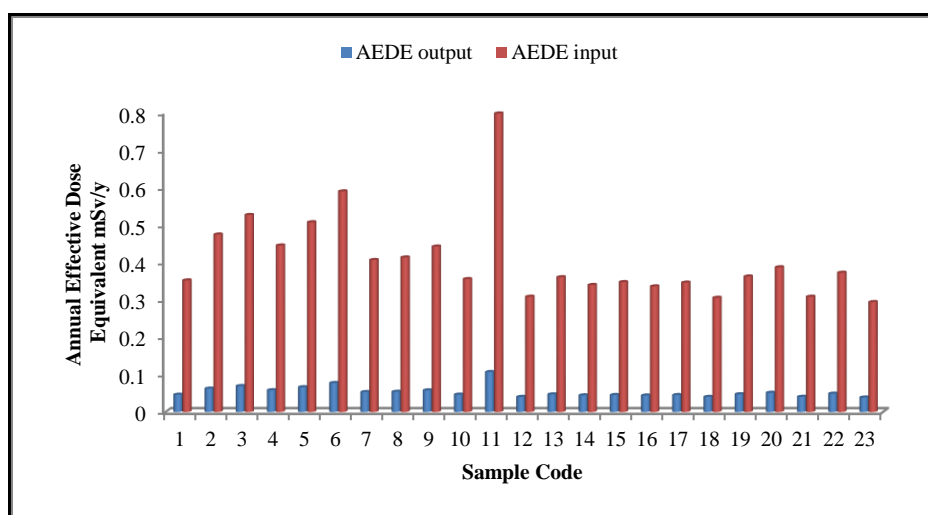


Fig. 6: Annual Effective Dose of sediment samples.

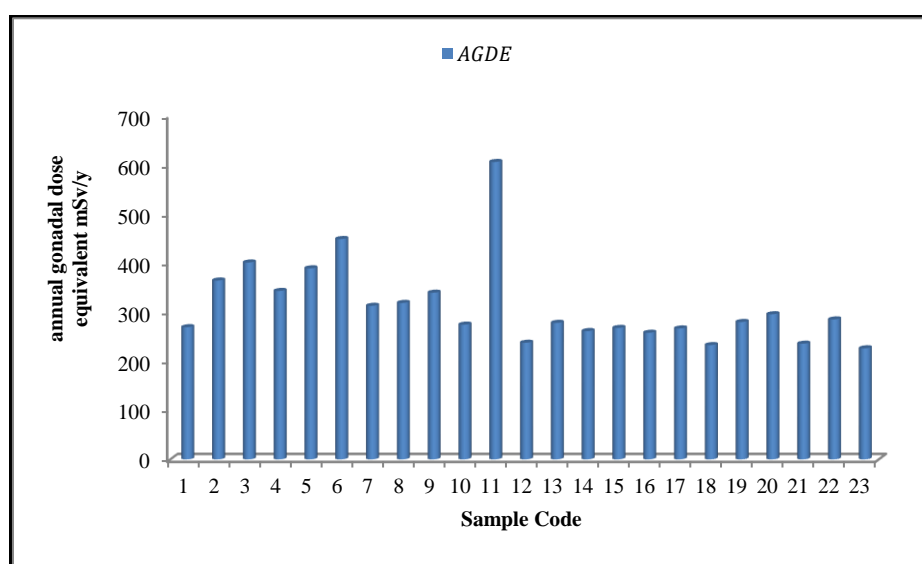


Fig. 7: Annual gonadal dose equivalent of sediment samples.

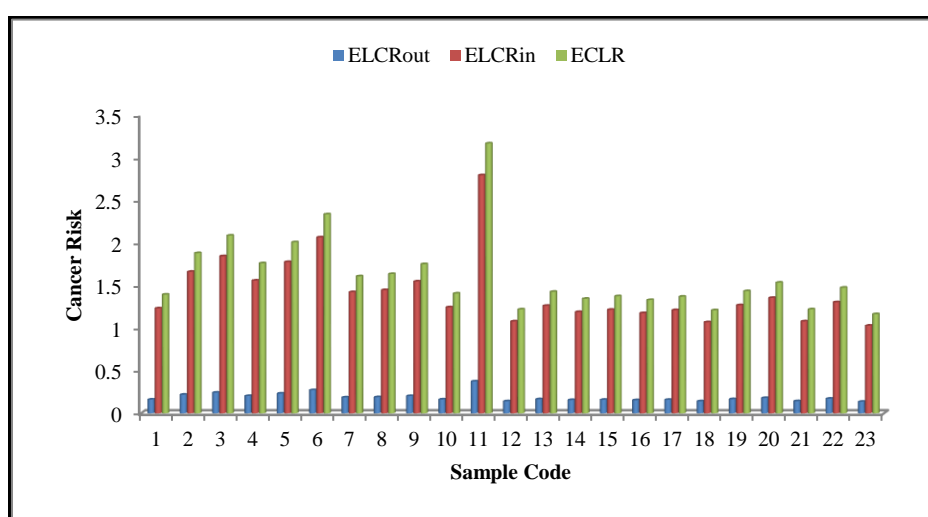


Fig. 8: Excess Lifetime Cancer Risk of Sediment Samples.

## 4. Conclusions

The activity concentrations of  $^{226}\text{Ra}$ ,  $^{232}\text{Th}$ , and  $^{40}\text{K}$  in the studied samples are found to be normal and below the average global values. The average radium equivalent activity ( $R_{\text{eq}}$ ) and external hazard ( $H_{\text{ex}}$ ) indices were less than the world average values. The calculated average absorbed dose rate and estimated, annual effective dose, annual gonadal dose equivalent (*AGDE*), excess life-time cancer risk (*ELCR*) were found lower than the worldwide average values for Marine sediments samples. By reference to the values of radiation hazard indices for all marine sediments samples collected from Coastal Marine of Abian beach, Aden Gulf, Yemen. Therefore, the probability of the radiological impact on the inhabitants/public living in this area will be insignificant. It is not also hazardous for tourist because they stay for short period in the concerning area. The radioactivity levels in sediments are a source of radiation exposure for marine organisms. Regular monitoring of radioactivity levels is vital for radiation risk confinement. The results provide important baseline data to which future radioactivity levels in marine environments can be compared. Considering the fact that oceans and seas form the ultimate sink of contaminants, including radioactivity, future research initiatives that study radioactivity levels in marine environments and assess associated radiological hazards to the population are of utmost importance in order to ensure protection of the marine environment. Such a project should also consider investigating radioactivity from artificial radionuclides.

## References

- [1] United Nations Scientific Committee on the Effects of Atomic Radiation, Report to the General Assembly, UNSCEAR, United Nations, New York, 1988.
- [2] United Nations Scientific Committee on the Effects of Atomic Radiation (UNSCEAR). Sources, Effects and Risks of Ionization Radiation; Report to the General Assembly, with Scientific Annexes B: Exposures from Natural Radiation Sources; UNSCEAR: New York, NY, USA, 1993.
- [3] M. Iqbal, M. Tufail and S.M. Mirza, "Measurement of Natural Radioactivity in Marble Found In Pakistan Using a NaI(Tl) Gamma-Ray Spectrometer". Technical Note, Journal of Environmental Radioactivity, 51(2), 255–265, 2000.
- [4] A. Abu Taleb, A. Abbady and S Harb, Assessment of Natural Radioactivity Level in Shore Sediment Samples from Nasser Lake at Aswan, Egypt. (IJBS), Vol. 6, No. 1, January 2019.
- [5] Bathan, K., Mehra, R., Sonkawade, R. G., Singh, S. "Use of Gamma-Ray Spectrometry for Assessment of Natural Radioactive Dose in Some Sample of Building Materials", Asian Journal of Chemistry, Vol. 21, No. 10, PP. 207-21, 2009.
- [6] Ahmed et al. "Natural Radioactivity and Dose Assessment in Sand and Sediment Samples from Kuakata Beach, Bangladesh". J. Bangladesh Acad. Sci., Vol. 40, No. 1, 45-55, 2016.
- [7] International Atomic Energy Agency (IAEA). Extent of Environmental Contamination by Naturally Occurring Radioactive Material (NORM) and Technological Options for Mitigation; IAEA Technical Reports Series No. 419; IAEA: Vienna, Austria, 2003.
- [8] I.H. Salah, Radioactivity of  $^{238}\text{U}$ ,  $^{232}\text{Th}$ ,  $^{40}\text{K}$ , and  $^{137}\text{Cs}$  and assessment of depleted uranium in soil of the Musandam Peninsula, Sultanate of Oman. Turkish J. Eng. Environ. Sci. 36, 236 – 248. 2012.
- [9] UNSCEAR, United Nations Scientific Committee on the Effects of Atomic Radiation. Report to the General Assembly United Nations, New York. Annex B, Vol. I, 223–439, 2008.
- [10] J. Al-Jundi, B.A. Al-Bataina, Y. Abu-Rukah, H.M. Shehadeh, Natural radioactivity concentrations in soil samples along the Amman Aqaba Highway, Jordan. Radiat. Measur. 36, 555–560, 2003.
- [11] H. Orabi, A. Al-Shareaif, M. El Galefi, Gamma-ray measurements of naturally occurring radioactive sample from Alkharje city. J. Radioanal. Nucl. Chem. 269, 99–102, 2006.
- [12] A.A. Abdel-Halim, I.H. Saleh, Radiological characterization of beach sediments along the Alexandria–Rosetta coasts of Egypt. Journal of Taibah University for Science 10(2016) 212–220. 2016.
- [13] N.B. Baggoura, Plutonium isotopes,  $^{137}\text{Cs}$ ,  $^{90}\text{Sr}$  and natural radioactivity in marine sediments from Ghazaouet (Algeria), J. Environ. Radioact. 34 (2) 127–138, 1997
- [14] ICRP. The 2007 Recommendations of the International Commission on Radiological Protection (ICRP); ICRP Publication 103; Ann. ICRP 37; Pergamon Press: Oxford, UK, 2007.
- [15] IAEA. Radiation Protection and Safety of Radiation Sources: International Basic Safety Standards; International Atomic Energy Agency (IAEA): Vienna, Austria, 2014.

- [16] I. I. Suliman and K. Alsafi, "Radiological Risk to Human and Non-Human Biota Due to Radioactivity in Coastal Sand and Marine Sediments, Gulf of Oman," *Life*, vol. 11, no. 6, p. 549, Jun. 2021, doi: <https://doi.org/10.3390/life11060549>.
- [17] J. Valentin, *Environmental Protection: The Concept and Use of Reference Animals and Plants*. Annals of the ICRP; ICRP Publication 108; ICRP: Ottawa, ON, Canada, 2008.
- [18] M. Q. Jaber et al., The concentrations of radon in the marine sediments of Ra's Al-Besha, Northern west of the Arabian Gulf, *Mesopot. J. Mar. Sci.*, 32 (1): 1–8, 2017.
- [19] N.M. Alam, M.I. Chowdury, M. Kamal, S. Ghose, N.M. Islam, M.N. Mustafa, M.M. Miah, M.M. Ansary, The  $^{226}\text{Ra}$ ,  $^{232}\text{Th}$  and  $^{40}\text{K}$  activities in beach sand minerals and beach soils of Cox's Bazar (Bangladesh), *J. Environ. Radioact.* 46 (1999) 243–261, 1999.
- [20] UNSCEAR 2012 Effects and Risks of Ionizing Radiation", United Nations Scientific Committee on the Effects of Atomic Radiation Report to the General Assembly, with Scientific Annexes A and B Series. Annex A - Attributing health effects to ionizing radiation exposure and inferring risks. Annex B-Uncertainties in risk estimates for radiation-induced cancer. United Nations New York.
- [21] I. Al-Qaradawi, M. Abdel-Moati, M.A.A. Al-Yafei, E. Al-Ansari, I. Al-Maslamani, E. Holm, I. Al-Shaikh, A. Muring, P.V. Pinto, D. Abdulmalik, et al., Radioactivity levels in the marine environment along the Exclusive Economic Zone (EEZ) of Qatar. *Mar. Pollut. Bull.*, 90, 323–329, 2015.
- [22] S. Uddin, A. Aba, S.W. Fowler, M. Behbehani, A. Ismaeel, H. Al-Shammari, A. Alboloushi, J.W. Mieltski, A. Al-Ghadban, A. Al-Ghunaim, et al., Radioactivity in the Kuwait marine environment-Baseline measurements and review. *Mar. Pollut. Bull.*, 100, 651–661, 2015.
- [23] A.Z. Al-Zamel, F. Bou-Rabee, M. Olszewski, H. Bem, Natural radionuclides and  $^{137}\text{Cs}$  activity concentration in the bottom sediment cores from Kuwait Bay. *J. Radioanal. Nucl. Chem.*, 266, 269–276, 2005.
- [24] H.A. Al-Trabulsy, A.E.M. Khater, F.I. Habbani, Radioactivity levels and radiological hazard indices at the Saudi coastline of the Gulf of Aqaba. *Radiat. Phys. Chem.*, 80, 343–348, 2011.
- [25] X. Lu, and X. Zang, Measurement of natural radioactivity in beach sands from Rizhao bathing beach, China. *Radiation protection dosimetry*, 130, 385–388, 2008.
- [26] Y. Örgün, N. Altinsoy, S.Y. Sahin, Y. Güngör, A.H. Gültekin, G. Karahan, and Z. Karacik, Natural and anthropogenic radionuclides in rocks and beach sands from Ezine region (Çanakkale), Western Anatolia, Turkey. *Applied Radiation and Isotopes*, 65, 739–747, 2007.
- [27] M. Casas-Ruiz, R.A. Ligeró, and L. Barbero, Estimation of annual effective dose due to natural and man-made radionuclides in the metropolitan area of the Bay of Cadiz (SW of Spain). *Radiation protection dosimetry*, 15, 60–70, 2012.
- [28] S.H.Q. Hamidaddin, Measurements of the natural radioactivity along Red Sea coast (South beach of Jeddah Saudi Arabia). *Life Sci. J.* 10 (1), 121–128, 2013.
- [29] Fatimh Alshahri. Radioactivity of  $^{226}\text{Ra}$ ,  $^{232}\text{Th}$ ,  $^{40}\text{K}$  and  $^{137}\text{Cs}$  in beach sand and sediment near to desalination plant in eastern Saudi Arabia: Assessment of radiological impacts, *Journal of King Saud University – Science* (2017) 29, 174–181, 2017.
- [30] J. Wang, J. Du, Q. Bi, Natural radioactivity assessment of surface sediments in the Yangtze Estuary. *Mar. Pollut. Bull.*, 114, 602–608, 2017
- [31] J. Beretka, P.J. Mathew, Natural radioactivity of Australian building materials, industrial wastes and by-products. *Health Phys.*, 48, 87–95, 1985.
- [32] E.I. Hamilton, The relative radioactivity of building materials. *Am. Ind. Hyg. Assoc. J.*, 32, 398–403, 1971.
- [33] UNSCEAR 2000 United Nations Scientific Committee on the Effects of Atomic Radiation. Sources and Effects of Ionizing Radiation. Report to the General Assembly, with Scientific Annexes Volume I: Sources, United Nations, New York.
- [34] UC 1999 European Commission. Radiological protection principles concerning the natural radioactivity of building materials In EC radiation Protection 112 Directorate General Environment, Nuclear Safety and Civil Protection.
- [35] V. Ramasamy, M. Sundarajan, G. Suresh, K. Paramasivam and V. Meen akshisundaram, Role of light and heavy minerals on natural radioactivity level of high background radiation area, Kerala, India *Appl. Radiat. and Isotopes* 85 110, 2014. <http://dx.doi.org/10.1016/j.apradiso.2013.11.119>.

- [36] I. T. AL-Alawy, R. S. Mohammed, H. R. Fadhil and A. A. Hasan. "Determination of Radioactivity Levels, Hazard, Cancer Risk and Radon Concentrations of Water and Sediment Samples in Al-Husseiniya River (Karbala, Iraq)", IOP Conf. Series: Journal of Physics: Conf. Series 1032 012012 2018, doi: 10.1088/1742-6596/1032/1/012012.
- [37] ICRP, International Commission on Radiological Protection). Publication 119: Compendium of dose coefficients based on ICRP Publication 60. Annals of the ICRP 41 (suppl) 42(4) 1-130, 2012.

## تقدير مستوى الإشعاع الطبيعي والمخاطر المصاحبة للإشعاع لعينات رملية من بقايا البحر جمعت من شاطئ ساحل ابين على خليج عدن - اليمن

فؤاد عبده الصبيحي<sup>1\*</sup>، طاهر عبدالله الجبار سالم<sup>2</sup> و محمود عيسى احمد<sup>3</sup>

<sup>1</sup> قسم الفيزياء، كلية التربية - صبر، جامعة عدن، اليمن

<sup>2</sup> قسم الكيمياء، كلية التربية - صبر، جامعة عدن، اليمن

<sup>3</sup> قسم الفيزياء، كلية التربية - زنجبار، جامعة ابين، اليمن

\* الباحث الممثل: فؤاد عبده الصبيحي؛ البريد الإلكتروني: fabdo11@yahoo.com

استلم في: 20 ديسمبر 2023 / قبل في: 07 يناير 2023 / نشر في: 31 مارس 2023

### الملخص

تمّ تجميع 23 عينة بقايا بحرية من ساحل ابين على شاطئ خليج عدن لقياس تركيز الراديوم-226، الثوريوم-232 والپوتاسيوم-40 ومعاملات الاخطار الناتجة عنها باستخدام كاشف الجرمانيوم عالي النقاوة HPGe. اظهرت النتائج ان اعلى تركيز لفعالية نويدة الراديوم-226 وجدت في العينة S11 والذي تساوي  $(36.300 \pm 1.41 \text{ Bq/kg})$  بينما اقل قيمة وجدت في العينة S23 والذي تساوي  $(9.44 \pm 0.38 \text{ Bq/kg})$  بمعدل تركيز يساوي  $(16.25 \pm 0.53 \text{ Bq/kg})$ . بينما وجدت اعلى تركيز لفعالية نويدة الثوريوم-232 المشعة طبيعياً في العينة S11 والذي تساوي  $(95.75 \pm 5.88 \text{ Bq/kg})$  واقل قيمة وجدت في العينة S12 والذي تساوي  $(12.94 \pm 0.79 \text{ Bq/kg})$  بمعدل تركيز يساوي  $(23.80 \pm 1.46 \text{ Bq/kg})$ . اما اعلى تركيز لفعالية نويدة البوتاسيوم-40 فقد وجدت في العينة S5 والتي تساوي  $(747.72 \pm 64.32 \text{ Bq/kg})$  واقل قيمة وجدت في العينة S11 والذي تساوي  $(296.67 \pm 25.52 \text{ Bq/kg})$  بمعدل تركيز يساوي  $(518.54 \pm 45.84 \text{ Bq/kg})$ . تم حساب معاملات الاخطار الاشعاعية الناتجة عن تواجد هذه الانوية الطبيعية في عينات البقايا البحرية المقاسة. اظهرت النتائج ان فعالية المعامل الراديومي ( $R_{eq}$ ) تقع في المدى  $(65.58 \text{ to } 196.06 \text{ Bq/kg})$  بمعدل يساوي  $(90.21 \text{ Bq/kg})$ ، معامل الخطورة الخارجي ( $H_{ex}$ ) يقع في المدى  $(0.177 \text{ to } 0.529)$  بمعدل يساوي  $0.243$ ، معامل الخطورة الداخلي ( $H_{in}$ ) يقع في المدى  $(0.203 \text{ to } 0.627)$  بمعدل يساوي  $0.287$ . حسب معدل الجرعة الممتصة الخارجي  $D_{out}$  ومعدل الجرعة الممتصة الداخلي  $D_{in}$  ووجدت انها تقع في المدى  $(31.57 \text{ to } 87.06 \text{ nGy} \cdot \text{h}^{-1})$  بمعدل يساوي  $(59.91 \text{ to } 162.75 \text{ nGy} \cdot \text{h}^{-1})$  بمعدل يساوي  $43.66 \text{ nGy} \cdot \text{h}^{-1}$ ،  $83.13 \text{ nGy} \cdot \text{h}^{-1}$  على التوالي. حسبت الجرعة السنوية المؤثرة الخارجية ( $AED_{out}$ ) والداخلية ( $AED_{in}$ ) ووجدت انها تقع في المدى  $(0.039 \text{ to } 0.294 \text{ mSv} \cdot \text{h}^{-1})$  بمعدل يساوي  $0.107 \text{ mSv} \cdot \text{h}^{-1}$  و  $0.053 \text{ mSv} \cdot \text{h}^{-1}$  على التوالي. حسبت قيم الجرعة المكافئة السنوية للغة التناسلية AGDE ووجدت انها تقع في المدى  $(226.13 \text{ to } 605.55 \mu\text{Sv} \cdot \text{h}^{-1})$  بمعدل يساوي  $312.52 \mu\text{Sv} \cdot \text{h}^{-1}$ . بينما حسبت قيمة كل من (excess life-time cancer risk ( $ELCR_{out}$ ) cancer risk الخارجي وكذلك ( $ELCR_{in}$ ) excess life-time cancer risk الداخلي ووجدت انها تقع في المدى  $(0.135 \times 10^{-3} \text{ to } 0.374 \times 10^{-3})$  بمعدل يساوي  $(0.187 \times 10^{-3})$ ،  $(1.427 \times 10^{-3})$  على التوالي. جميع النتائج التي حصلنا عليها للعينات قيد الدراسة اقل من القيمة العليا المسموح بها والموصى بها من The World Health Organization and EU Council وان الاشعاعات الناتجة عن النويدات الطبيعية لا تشكل خطورة على الحياة البشرية.

**الكلمات المفتاحية:** النشاط الإشعاعي الطبيعي، ساحل ابين، اخطار الاشعاعات الطبيعية، البقايا البحرية، كاشف الجرمانيوم عالي النقاوة.

### How to cite this article:

F. A. AS-Subaihi, T. A. A. Salem and M. I. Ahmed "ASSESSMENT OF NATURAL RADIOACTIVITY LEVEL AND ASSOCIATED RADIOLOGICAL HAZARDS IN MARINE SEDIMENT SAMPLES COLLECTED FROM ABYAN BEACH, GULF OF ADEN, YEMEN", *Electron. J. Univ. Aden Basic Appl. Sci.*, vol. 4, no. 1, pp. 18-30, Mar. 2023. DOI: <https://doi.org/10.47372/ejua-ba.2023.1.217>



Copyright © 2023 by the Author(s). Licensee EJUA, Aden, Yemen. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY-NC 4.0) license.